

# EFFECT OF DUMPSITE WASTE ON SOIL, DRINKING WATER SOURCES, AND ANTIBIOTIC SUSCEPTIBILITY OF ISOLATES, IN A SUB-URBAN COMMUNITY, BENIN CITY, NIGERIA

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## ABSTRACT

Poor management of dumpsite waste in semi-urban communities could impact the soil and underground source of drinking water, thereby posing a health risk to residents. In this study, the effect of dumpsite waste, Evbotubu Quarters, Benin City, on the physicochemical properties, heavy metals, and minerals content of soil, well water and borehole water, antibiotic susceptibility of isolates from the samples were determined. The frequency of occurrence of organisms isolated include *Bacillus cereus* (22.4%), *Escherichia coli* (12.2%), *Enterococcus* spp. (12.2%), *Staphylococcus aureus* (10.2%), *Pseudomonas aeruginosa* (8.2 %), *Streptococcus* spp. (8.2 %), *Corynebacterium* spp. (8.2 %), *Aspergillus* spp. (8.2 %), *Fusarium* spp. (6.2 %), and *Mucor* spp. (4 %). Although, the Ca, Mg, Na, Zn, Cu, Cr, Pb, Mn, and Fe content of the dumpsite soil were higher than other samples, there are some exceptions with regards to the physicochemical parameters. The borehole water had the lowest values for all the parameters, except pH (6.87±0.03). The results reported were within the World Health Organization (WHO) permissible limits, with few exceptions. Antibiotic susceptibility test showed that ciprofloxacin was the most effective antibiotic. Going by the level of contamination, relevant government agencies should strictly monitor waste disposal, ensure that boreholes, and wells are situated very far from dumpsites.

**Keywords:** Antibiotic resistance, bioaccumulation, heavy metals, waste management, waterborne diseases.

## INTRODUCTION

In developing countries, suburban communities are responsible for solving their water needs with little or no assistance from the government. The communities rely on ground water as the safest source of drinking water (Ishaku *et al.*, 2011). The contamination of various water sources could render it unsafe for drinking, unless it undergoes adequate treatment (Nwiniyi *et al.*, 2020). Going by the economic reality in Nigeria today, proper treatment of water before drinking is expensive for many inhabitants of suburban communities. Majority of the people are left with no choice than to drink untreated water that predisposes them to high risk of waterborne diseases such as cholera, typhoid, salmonellosis, *Escherichia coli* diarrhea, among others (Forstintus *et al.*, 2016; Manetu and Karanja, 2021).

Waste generated by households, markets, schools, small scale industries, etc. in suburban communities are usually transported

using garbage trucks to landfill or open dumpsites. Waste scavengers visit the dumpsites to sort waste, and sell the ones that could be recycled. Wastes dumped at landfills are subject to either groundwater infiltration from precipitations or erosion which usually contaminates surface water (Amano *et al.*, 2021). Improper waste disposal practices especially improper management of dumpsites in sub-urban communities generate leachate that often contaminate the soil surface and underground water (Olukanni *et al.*, 2017; Yahaya *et al.*, 2022a). Toxic leachates from the dumpsite wastes are capable of impacting the environment. Leachates are most likely to migrate to the water bodies, underground water, soils, and other biophysical components of the environment, resulting in adverse effect on humans, aquatic organisms, plants, and animals (Agbeshiea *et al.*, 2020). According to Bhalla *et al.* (2012), leachates can permeate the soil and ground water close to landfill sites. The resulting effect is pollution of subsoil and ground water through a combination of chemical, physical, and microbial processes in the dumpsite (Kjeldsen *et al.*, 2010). The exposure of groundwater to leachate is further increased by excess rainwater (Nagarajan *et al.*, 2012).

The indiscriminate disposal of waste, burning or decay of dumpsite wastes accumulates obnoxious materials that contain toxic heavy metals (Abdus-Salam, 2009). Pb, Hg, and Cd found in the dumpsites, and its underlying soils could easily be absorbed by plants and other soil microorganisms. Consequently, these heavy metals could enter the food chain, and pose a health risk to humans (Akanchise *et al.*, 2020; Nkwunonwo *et al.*, 2020). Waste dumpsites with a high level of heavy metals are most likely to affect the chemical properties of soil and its nutrients, which ultimately affect crop production, and diversity of soil microorganisms (Fosu-Mensah *et al.*, 2017).

In the last few years, Nigerian urban population has witnessed a remarkable growth due to rural-urban migration (Oyeleye, 2013). High population density in urban and semi-urban areas, especially in developing countries like Nigeria, are associated with poor handling of waste materials of all kinds generated by the inhabitants and few industries (Adekola and Ogundipe, 2017; Enerijiofi and Ekhaise, 2019; Ogbu and Ezeodili, 2021). This problem is compounded by inadequate modern waste disposal facilities, poor physical planning of semi-urban areas, and bad habit of individuals who litter the environment with waste. Consequently, the environment becomes highly polluted, with its attendant health risk to the residents (Adejobi and Olorunnimbe, 2012).

According to Atikpo *et al.* (2022), the average generation rate of solid waste in Evbotubu community, Edo State, Nigeria is 0.45 kg/p/d. The highest waste generation rate was recorded within the period of December to April. Although the result is within the range of 0.4-0.6 kg/p/d recommended by the United Nations Environmental Program for sub-urban communities in developing countries (Joarder, 2000), a lot of houses in Evbotubu community are barely completed with underground boreholes. Most of the boreholes have no overhead tanks. What is commonly seen in the locality are hand-dug wells close to open dumpsites. Indiscriminate disposal of solid and semi-solid waste in open dumpsites close to water sources has created serious environmental degradation problems translating into economic, social, and health issues (Ukpong *et al.*, 2015). All kinds of waste in the dumpsites are usually left for several days and weeks before waste collectors arrive to evacuate it. Unlawful dumping of municipal solid waste in unapproved sites comes with grave health consequences (Tukura *et al.*, 2018; Ichipi and Senekane, 2023). Waste that accumulate in the dumpsites promote the proliferation of pathogenic microorganisms, vectors, flies, insects, and rodents of public health importance (Ahmed, 2011; Nor Faiza *et al.*, 2019). Oyekan and Sulyman (2013) reported that the insects and vectors that transmit diseases of public health importance are usually found in the dumpsites. Vegetation grown in contaminated soil has the ability to bioaccumulate toxic heavy metals in their tissues that pose potential health risks, as they could biomagnify in the consumer's tissues (Onwughara *et al.*, 2010).

Although several studies have been carried out to evaluate the quality of water obtained from boreholes and wells located near dumpsites (Olukanni *et al.*, 2017; Ugbebor and Ntesat, 2019; Didia and Weje, 2020; Yahaya *et al.*, 2022a, Yahaya *et al.*, 2022b; Ugwoha *et al.*, 2023), there is limited information on the microbial species detected in the water samples collected from different dumpsites and antibiotic susceptibility pattern of the isolates. Therefore, this study is aimed at evaluating the effect of dumpsite waste, Evbotubu quarters, Benin City, on selected physicochemical parameters, heavy metals, minerals content, and antibiotic susceptibility of isolates from dumpsite soil, well water, and borehole water.

## MATERIALS AND METHODS

### Study area

The study was conducted at the Evbotubu, in Egor Local Government Area of Edo State, Nigeria. Evbotubu is a sub-urban area in Benin City, located geographically at latitude 6° 20' 3.3" N and longitude 5° 36' 50.724" E.

### Sample collection

Soil and water samples were collected in triplicates from the waste dumpsites. Soil samples were collected from the soil beneath the waste dump at depth 0-5 cm (Sample A); soil leachate (Sample B) using a hand trowel cleaned with 70% alcohol. Both samples were poured inside polyethylene bags. Water (Sample C) from a well close to the dumpsite, and sample D (borehole water in the vicinity of the dumpsite) were collected using a plastic screw capped sterilized bottles (2 litre capacity). All the samples were quickly taken to the laboratory within 2 hours of collection for analysis (Oviasogie *et al.*, 2010; Obire *et al.*, 2002).

### Serial dilution

One millilitre (1 mL) of water and one gram (1 g) soil sample each was weighed and mixed properly in 9 mL distilled water. Then, 1 mL of the dilution was transferred aseptically into another test tube containing 9 mL distilled water, and diluted serially into other test tubes till a  $10^{-5}$  dilution was achieved using a sterile pipette for each transfer.

### Microbiological analyses

#### Determination of heterotrophic bacterial count and total fungal count

An aliquot of 0.1 mL of each dilution was aseptically plated out using pour plate method on tryptone soy blood agar (Lab M Ltd UK), deoxycholate citrate agar (Lab M Ltd, UK), and nutrient agar (Lab M Ltd UK) for bacteria isolation, while potato dextrose agar (Lab M, UK) supplanted with 0.5% chloramphenicol is for fungi. Isolation and enumeration of the total heterotrophic bacterial count (THBC), and total fungal count (TFC) was done in triplicates. All culture plates were incubated at 37 °C for 24 to 48 hour to obtain bacterial counts; at 25 °C for 72 hours for fungal counts. The discrete colonies on the different agar media were observed, and counted in a Techmel and Techmel counter model TT 201. The results were expressed accordingly as colony forming units per gram (CFU/g) and colony forming units per millilitre (CFU/mL).

#### Determination of pure isolates

Different bacterial colonies that appeared on the culture plates were subcultured onto freshly prepared nutrient agar plates, and incubated at 28 °C for 24 hours. This procedure was repeated until discrete colonies were obtained in the culture plates. The discrete colonies were transferred aseptically onto nutrient agar slants, and incubated at 28 °C for 24 hours.

#### Characterization of bacterial and fungal isolates

Representatives of the pure bacterial colonies were subjected to various cultural, morphological, and biochemical tests. Identification of the isolates was based on Bergey's Manual of Determinative Bacteriology (Holt *et al.*, 2000). For the fungal isolates, wet mount method was used to determine their macroscopic and microscopic characteristics (Tsuneo, 2010).

#### Heavy metals, minerals and physicochemical analyses

The minerals content and physicochemical parameters of the soil and water samples were determined using the standard methods. Quantitative analysis of minerals (calcium, magnesium, and manganese), and heavy metals (copper, zinc, chromium, iron, and lead) in the dumpsite soil, dumpsite leachate, well water, and borehole water samples were determined using atomic absorption spectrophotometer (AAS), and the procedure described by the Association of Official Analytical Chemists (AOAC, 2000). The potassium and sodium contents of the samples were determined using flame emission spectrophotometer (AOAC, 2000). The pH of soil and water samples was determined using a Hach pH meter, in line with the procedure described by the American Public Health Association (APHA, 2005). The conductivity test was carried out using a Hach Conductivity meter (APHA, 2005). The total organic carbon in the samples was determined by the method of ASTM (2003). The alkalinity, sulphate, nitrate, phosphate, biochemical oxygen demand (BOD), total suspended solids (TSS), total

dissolved solids (TDS), and chloride concentrations in the soil and water samples were determined using the method of APHA (2005). The concentration of soil in the suspension that determines the level of organic matter, silt, clay, and sand in the samples were carried out using a simplified hydrometer procedure described by Moorberg and Crouse (2021).

#### Antimicrobial susceptibility test

The Kirby-Bauer disk diffusion method described by Jorgensen and Turnidge (2007), was used to evaluate the antimicrobial susceptibility of the bacterial isolates from the dumpsite soil, dumpsite leachate, well water, and borehole water samples. The commercially prepared antibiotics used include Amoxicillin (10 µg), Cotrimoxazole (30 µg), Peflacin (30 µg), Ofloxacin (5 µg), Augmentin (30 µg), Nitrofurantoin (300 µg), Tetracycline (30 µg), Gentamicin (10 µg), Erythromycin (30 µg), Streptomycin (30 µg), Cefazidime (30 µg), and Ciprofloxacin (5 µg). Overnight broth cultures of pure bacterial and fungal isolates were suspended in sterile peptone water. Using a sterile swab stick, each isolate was inoculated into the surface of dried plates of Mueller-Hinton agar, and rotated carefully to ensure even distribution (Salaimon *et al.*, 2015). The surface of the inoculated plates was allowed to dry for few minutes. Thereafter, a sterile forceps was used to lift antibiotic discs from the manufacturer's pack and placed on the inoculated plates. The plates were incubated at 37 °C for 24 hours. The diameter of the zone of inhibition of each antibiotic in the disk was measured to the nearest whole number, and interpreted using the guidelines of the Clinical and Laboratory Standard Institute (CLSI, 2005).

#### Statistical analysis

The results were presented as mean standard values of triplicate analysis. A one-way analysis of variance (ANOVA) and student t-test was performed using the data generated from the analyses as described by Ogbelibe (2005). Duncan's Multiple Range Test (DMRT) was carried out to test for significant difference at  $P < 0.05$  between treatment combinations. Statistical Package for the Social Sciences (SPSS) version 16 software was used for easy analysis of data.

## RESULTS AND DISCUSSION

The result of biochemical and physicochemical properties of dumpsite soil (sample A), dumpsite leachate (sample B), well water (sample C), and borehole water (sample D) are presented in Table 1. The pH values range from  $5.26 \pm 0.03$  to  $6.87 \pm 0.04$ . The pH permissible limit stipulated by the World Health Organization (WHO, 2010) is 6.5-8.5. Among all the four samples analyzed for pH, sample C and sample D were within the WHO permissible limit. The distance both water sources maintained from the dumpsite could be one of the reasons the water samples was slightly acidic (6.45-6.87). The pH of dumpsite soil (sample A) and dumpsite leachate (sample B) was moderately acidic (5.26-5.71), but failed to meet the WHO requirement. This could be attributed to the nature of waste materials dumped on the ground, chemical processes, breakdown, and bioaccumulation of organic waste that infiltrate underground water, and acidic rainwater. According to Ano and Okwunodudu (2008), the acidic nature of soil samples in the waste sites could leach to the underground water and affect its

taste.

The pH of a sample is positively correlated with electrical conductivity and total alkalinity (Gupta and Shukla, 2006). The electrical conductivity (EC) values of dumpsite soil, dumpsite leachate, well water, and borehole water samples range from  $15.00 \pm 0.08$  to  $132.00 \pm 1.52$  µS/cm. The dumpsite leachate sample B had the highest EC ( $132.00 \pm 1.52$  µS/cm). The presence of ionic contaminants, chemicals, and salts from all kinds of wastes dumped on the ground could be responsible for high EC of dumpsite leachate. The electrical conductivity (EC) values of all the samples were below the WHO permissible limit.

The total suspended solids (TSS) and total dissolved solids (TDS) in the water samples range from  $0.06 \pm 0.00$ - $23.24 \pm 0.30$  mg/L and  $7.00 \pm 0.01$ - $55.50 \pm 0.01$  mg/L, respectively. Higher values of TSS and TDS in the well water sample, compare to the dumpsite soil, dumpsite leachate, and borehole water could be attributed to its liquid state that dissolve or suspend solids, and exposure of the well water to contamination by foreign materials. For example, the runoff from the dumpsite wastes can enter a nearby well, and contaminate the water. According to Adejuwon and Mbuk (2011), high level of dissolved and suspended solids in water is caused by pollution, especially the presence of chemicals in the water sources. The biochemical oxygen demand (BOD) estimates the level of organic pollution of water. In the dumpsite soil and dumpsite leachate, BOD was not detected. However, the BOD values of water samples C and D was  $4.35 \pm 0.41$  and  $1.67 \pm 0.10$  mg/L, respectively. According to the classification of surface water quality, BOD > 2.0 mg/L is indicative of pollution of the water source, and regarded as unfit for human consumption (Afolabi *et al.*, 2012). However, the BOD of the water samples was below the WHO permissible limit (6.0 mg/L).

The chloride, sulphate, and nitrate content of the dumpsite soil, dumpsite leachate, well water, and borehole water samples were within the WHO permissible limits. A similar result was reported for the phosphate content of all the samples, with exception of the well water sample. The excess quantity of phosphate in the well water, could promote the growth of algae and production of slime, which compromises the quality of water (Ibiam *et al.*, 2024). Although, the WHO Standard for alkalinity of drinking water is yet to be ascertained, the Indian authority recommend that the water for domestic use, and drinking water should have values below 200 ppm (Akhtar *et al.*, 2013). According to George *et al.* (2010), alkalinity influences the chemical and biochemical reactions in water. The total organic carbon (TOC), clay, silt, sand, and organic matter, were not detected in the well water and borehole water samples. However, different values were reported for each of the parameters in the dumpsite soil, and dumpsite leachate samples. The total organic carbon (TOC) reported in both samples were below the WHO permissible limit. The results obtained from this study substantially agrees with the findings by Olukanni *et al.* (2017), that determined the effect of dumpsite leachates from a tertiary institution on the quality of groundwater.

Table 1. Biochemical and physicochemical compositions of dumpsite soil, dumpsite leachate, well water, and borehole water

| Parameter      | Unit  | Samples                 |                          |                         |                         | WHO Permissible Limits | References for the WHO Standard |
|----------------|-------|-------------------------|--------------------------|-------------------------|-------------------------|------------------------|---------------------------------|
|                |       | A                       | B                        | C                       | D                       |                        |                                 |
| pH             |       | 5.71±0.05 <sup>a</sup>  | 5.26±0.03 <sup>a</sup>   | 6.45±0.02 <sup>c</sup>  | 6.87±0.03 <sup>e</sup>  | 6.5-8.5                | Sargar et al. (2024)            |
| EC             | µS/cm | 95.00±0.13 <sup>a</sup> | 132.00±1.47 <sup>d</sup> | 41.00±1.90 <sup>b</sup> | 15.00±0.26 <sup>a</sup> | 500                    | Odu et al. (2021)               |
| Chloride       | mg/L  | 16.87±0.39 <sup>a</sup> | 23.76±0.79 <sup>d</sup>  | 9.98±0.14 <sup>b</sup>  | 8.59±0.17 <sup>a</sup>  | 200 mg/L               | Sargar et al. (2024)            |
| Sulphates      | mg/L  | 4.25±0.02 <sup>c</sup>  | 5.98±0.06 <sup>e</sup>   | 2.03±0.05 <sup>b</sup>  | 0.18±0.04 <sup>a</sup>  | 250-500 mg/L           | Odu et al. (2020)               |
| Phosphates     | mg/L  | 2.81±0.06 <sup>c</sup>  | 3.96±0.07 <sup>e</sup>   | 1.33±0.03 <sup>b</sup>  | 0.00±0.00 <sup>a</sup>  | 0.5 mg/L               | Odu et al. (2020)               |
| Nitrates       | mg/L  | 7.59±0.08 <sup>c</sup>  | 8.99 ±0.07 <sup>e</sup>  | 1.72±0.07 <sup>b</sup>  | 0.78±0.02 <sup>a</sup>  | 10 mg/L                | Odu et al. (2020)               |
| Alkalinity     | mg/L  | ND                      | ND                       | 21.87±0.05 <sup>b</sup> | 9.40±0.02 <sup>a</sup>  | -                      | Sargar et al. (2024)            |
| BOD            | mg/L  | ND                      | ND                       | 4.35±0.41 <sup>b</sup>  | 1.67±0.10 <sup>a</sup>  | 6.0 mg/L               | Oluokosi et al. (2016)          |
| TDS            | mg/L  | ND                      | ND                       | 55.50±0.01 <sup>b</sup> | 7.00±0.01 <sup>a</sup>  | 1000                   | Akhtar et al. (2014)            |
| TSS            | mg/L  | ND                      | ND                       | 23.24±0.30 <sup>b</sup> | 0.06±0.00 <sup>a</sup>  | 1000                   | Odu et al. (2020)               |
| TOC            | mg/L  | 3.32±0.01 <sup>b</sup>  | 0.53±0.01 <sup>a</sup>   | ND                      | ND                      | 50 mg/L                | Pandey et al. (2021)            |
| Organic matter | %     | 0.65±0.02 <sup>a</sup>  | 0.65±0.02 <sup>a</sup>   | ND                      | ND                      |                        |                                 |
| Sand           | %     | 71.98±0.04 <sup>b</sup> | 63.24±0.05 <sup>a</sup>  | ND                      | ND                      |                        |                                 |
| Silt           | %     | 17.81±0.03 <sup>a</sup> | 19.76±0.03 <sup>b</sup>  | ND                      | ND                      |                        |                                 |
| Clay           | %     | 17.00±0.04 <sup>b</sup> | 10.21±0.02 <sup>a</sup>  | ND                      | ND                      |                        |                                 |

Values represent a triplicate mean  $\pm$  standard error. The means with different superscripts in the same row are significantly different at  $p \leq 0.05$ .

Key: Sample A - dumpsite soil; Sample B - dumpsite leachate; Sample C - well water close to dumpsite; Sample D - borehole water in the vicinity of dumpsite; EC - Electrical conductivity; BOD - Biochemical oxygen demand; TDS - Total dissolved solids; TSS - Total suspended solids; TOC - Total organic carbon; ND - Not detected

Presented in Table 2 is the minerals and heavy metals content of dumpsite soil, dumpsite leachate, well water, and borehole water. The quantity of each of the minerals or heavy metals in the four samples is significantly different ( $p < 0.05$ ). It is worthy to note that the quantity of each mineral (Ca, Mg, Na, K, and Mn), and heavy metal (Cu, Cr, Fe, Zn, and Pb) in the dumpsite soil followed by dumpsite leachate was higher than the result reported in other samples. This observation could be attributed to human activities in the area, that generate a large quantity of household wastes, plumbing, gasoline, iron, industrial waste products etc. (Afolabi et al., 2012). On the contrary, borehole water in the vicinity of the dumpsite had the lowest quantity of the minerals and heavy metals. Calcium, potassium, sodium, magnesium, and zinc contents of all the samples were below the WHO permissible limits.

All the samples also contain copper and iron below the WHO permissible limit, with exception of dumpsite soil and dumpsite leachate. However, the chromium and lead content of all the samples were above the WHO permissible limit, with the exception of borehole water. According to Bhalla et al. (2012), heavy metals contamination of boreholes and hand dug wells could be attributed to infiltration or permeation of the landfill leachate through the

water. High toxicity, mutagenic, and carcinogenic properties of hexavalent chromium could be harmful to health. According to Odu et al. (2020), drinking water that contains excess quantity of chromium could cause allergic dermatitis and cancer. Severe damage to reproductive, peripheral and central nervous system, poor functioning of the kidney, cardiovascular system and joints, hampering the synthesis of hemoglobin could be attributed to lead poisoning. It causes mutagenic, teratogenic, and carcinogenic effect in the human body (Nwinyi et al., 2020). According to Udofia et al. (2016), lead poisoning could impair brain development and learning ability in children. Among all the samples, only the dumpsite soil contains manganese that was above the WHO permissible limit. The water from the well close to the waste dumpsite (Sample C) contain Cr (0.28±0.01 mg/L) and Pb (0.08±0.01 mg/L), that exceeded the WHO (2010) permissible limit. The implication is that the well water is unfit for human consumption. Drinking water that contains chromium above 0.05 mg/L can lead to cancer or allergic dermatitis. High concentration of lead in a water source is of public health significance because drinking the water might cause anaemia, kidney diseases, cancer, and affect vitamin D metabolism, and mental development in infants (NIS, 2007).

**Table 2.** Minerals and heavy metals content (mg/L) of dumpsite soil, dumpsite leachate, well water, and borehole water

| Parameters | Samples                 |                         |                        |                        | WHO Permissible Limit | References for the WHO Standard |
|------------|-------------------------|-------------------------|------------------------|------------------------|-----------------------|---------------------------------|
|            | A                       | B                       | C                      | D                      |                       |                                 |
| Calcium    | 6.86±0.05 <sup>d</sup>  | 5.77±0.04 <sup>c</sup>  | 1.87±0.04 <sup>b</sup> | 0.36±0.04 <sup>a</sup> | 75 mg/L               | Hossin et al. (2016)            |
| Magnesium  | 9.33±0.08 <sup>d</sup>  | 7.84±0.04 <sup>c</sup>  | 6.62±0.04 <sup>b</sup> | 1.02±0.03 <sup>a</sup> | 50 mg/L               | Hossin et al. (2016)            |
| Sodium     | 17.16±0.10 <sup>d</sup> | 12.18±0.10 <sup>c</sup> | 4.10±0.10 <sup>b</sup> | 0.43±0.05 <sup>a</sup> | 200 mg/L              | Hossin et al. (2016)            |
| Potassium  | 9.11±0.05 <sup>d</sup>  | 6.47±0.03 <sup>c</sup>  | 2.86±0.03 <sup>b</sup> | 0.67±0.01 <sup>a</sup> | 55 mg/L               | Hossin et al. (2016)            |
| Zinc       | 2.24±0.03 <sup>d</sup>  | 1.59±0.03 <sup>c</sup>  | 1.09±0.03 <sup>b</sup> | 0.81±0.04 <sup>a</sup> | 3.0 mg/L              | Mahugija (2018)                 |
| Copper     | 6.19±0.04 <sup>d</sup>  | 2.35±0.03 <sup>c</sup>  | 0.84±0.01 <sup>b</sup> | 0.00±0.00 <sup>a</sup> | 2.0 mg/L              | Manne et al. (2022)             |
| Chromium   | 7.40±0.05               | 0.33±0.01               | 0.28±0.01              | 0.00±0.00 <sup>a</sup> | 0.05 mg/L             | Khan et al. (2011)              |
| Lead       | 3.11±0.02 <sup>d</sup>  | 0.21±0.01 <sup>c</sup>  | 0.08±0.01 <sup>b</sup> | 0.00±0.00 <sup>a</sup> | 0.01mg/L              | Mahugija (2018)                 |
| Manganese  | 8.33±0.06 <sup>d</sup>  | 0.28±0.02 <sup>c</sup>  | 0.18±0.02 <sup>b</sup> | 0.12±0.01 <sup>a</sup> | 0.4 mg/L              | Obasi and Akudinobi (2020)      |
| Iron       | 28.86±0.15 <sup>d</sup> | 3.45±0.09 <sup>c</sup>  | 0.29±0.05 <sup>b</sup> | 0.07±0.03 <sup>a</sup> | 0.3 mg/L              | Mahugija (2018)                 |

Values represent a triplicate mean  $\pm$  standard error. The means with different superscripts in the same row are significantly different at  $p \leq 0.05$ .

Sample A - dumpsite soil; Sample B - dumpsite leachate; Sample C - well water close to dumpsite; Sample D - borehole water in the vicinity of dumpsite

Presented in Table 3 is the total heterotrophic bacterial and fungal count of dumpsite soil, dumpsite leachate, well water, and borehole water. The result shows that the Sample B (dumpsite leachate) had the highest bacterial and fungal counts of  $5.4 \times 10^5 \pm 0.1$  CFU/mL and  $2.1 \times 10^5 \pm 0.1$  CFU/mL, respectively. On the contrary, the borehole water in the vicinity of the dumpsite had the lowest heterotrophic bacterial and fungal count. In a related study, Ugbebor and Ntesat (2019) reported that the THBC ( $6.0 \times 10^3$  CFU/mL) of borehole water samples at Igwuruta solid waste dumpsite is the highest among all the samples. However, Yahaya et al. (2022a) and Yahaya et al. (2022b) reported that the total fungi/yeast in the water samples obtained from boreholes located near Olusosun dumpsite, and Simpson Transfer Loading Station in Lagos state is below the detection level. This result is not in agreement with the findings from this study.

Table 4 shows the distribution, frequency, and percentage occurrence of the bacterial isolates from the dumpsite soil, dumpsite leachate, well water, and borehole water. A total of seven bacterial species (*Bacillus cereus*, *Enterococcus* spp., *Streptococcus* spp., *Escherichia coli*, *Corynebacterium* spp., *Pseudomonas aeruginosa*, and *Staphylococcus aureus*), and three fungal species (*Fusarium* spp., *Aspergillus* spp., and *Mucor* spp.) were isolated from the samples. Among the isolates, *Bacillus cereus* and *Mucor* spp. had the highest and least percentage occurrence of 22.4 and 4.0%, respectively. This result is partly in agreement with a related research finding by Enerijiofi and Ekhaise (2019). The highest frequency of occurrence of *Bacillus cereus* in the samples could be attributed to its ability to form spores, well-known to be resistant to adverse environmental conditions, and ubiquitous nature of the bacterium (Ahaotu et al., 2021).

**Table 3.** Total heterotrophic bacterial and fungal count of dumpsite soil, dumpsite leachate, well water, and borehole water samples

| Samples    | Microbial Counts          |                           |
|------------|---------------------------|---------------------------|
|            | Bacterial counts          | Fungal counts             |
| A (CFU/g)  | $3.6 \times 10^5 \pm 0.1$ | $0.2 \times 10^5 \pm 0.1$ |
| B (CFU/g)  | $5.4 \times 10^5 \pm 0.1$ | $2.1 \times 10^5 \pm 0.1$ |
| C (CFU/mL) | $7.9 \times 10^3 \pm 0.2$ | $4.5 \times 10^2 \pm 0.2$ |
| D (CFU/mL) | $1.2 \times 10^1 \pm 0.1$ | $0.9 \times 10^1 \pm 0.1$ |

Key: Sample A - dumpsite soil; Sample B - dumpsite leachate; Sample C - well water close to dumpsite; Sample D - borehole water in the vicinity of dumpsite.

**Table 4.** Distribution, frequency, and percentage occurrence of bacterial and fungal isolates from the dumpsite soil, dumpsite leachate, well water, and borehole water samples

| Bacterial Isolates            | Sample    |           |           |           | Frequency | Mean percentage occurrence |
|-------------------------------|-----------|-----------|-----------|-----------|-----------|----------------------------|
|                               | A         | B         | C         | D         |           |                            |
| <i>Bacillus cereus</i>        | 4 (26.7)  | 4 (21.1)  | 2 (22.2)  | 1 (16.7)  | 11        | 22.4                       |
| <i>Enterococcus</i> spp.      | 3 (20.0)  | 2 (10.5)  | 1 (11.1)  | -(0.0)    | 6         | 12.2                       |
| <i>Streptococcus</i> spp.     | 1 (6.7)   | 2 (10.5)  | 1 (11.1)  | -(0.0)    | 4         | 8.2                        |
| <i>Escherichia coli</i>       | 2 (13.3)  | 3 (15.8)  | 1 (11.1)  | -(0.0)    | 6         | 12.2                       |
| <i>Corynebacterium</i> spp.   | 1 (6.7)   | 2 (10.5)  | 1 (11.1)  | -(0.0)    | 4         | 8.2                        |
| <i>Pseudomonas aeruginosa</i> | 1 (6.7)   | 2 (10.5)  | 1 (11.1)  | -(0.0)    | 4         | 8.2                        |
| <i>Staphylococcus aureus</i>  | 1 (6.7)   | 1 (5.3)   | 1 (11.1)  | 2 (33.3)  | 5         | 10.2                       |
| <i>Fusarium</i> spp.          | 1 (6.7)   | 1 (5.3)   | 1 (11.1)  | -(0.0)    | 3         | 6.2                        |
| <i>Aspergillus</i> spp.       | -(0.0)    | 1 (5.3)   | -(0.0)    | 3 (0.50)  | 4         | 8.2                        |
| <i>Mucor</i> spp.             | 1 (6.7)   | 1 (5.3)   | -(0.0)    | -(0.0)    | 2         | 4.0                        |
| Total number of isolates      | 15 (30.6) | 19 (38.8) | 09 (18.4) | 06 (12.2) | 49        | 100.0                      |

Key: Sample A = dumpsite soil; Sample B = dumpsite leachate; Sample C = well water close to dumpsite; Sample D = borehole water in the vicinity of dumpsite

The waste accumulated in the dumpsite is a reservoir of pathogenic infectious agents poses a great danger to public health. The high prevalence of *Escherichia coli* and *Enterococcus* spp., 12.2% each, in all the soil and water samples, except the sample D representing borehole water in the vicinity of the dumpsite, is an indication that the human and animal excreta is part of the waste in the refuse dumpsite. This is capable of causing outbreak of food and waterborne diseases (Adeyeba and Akinbo, 2002). The presence of these organisms in the well water, sample C indicates that the water is polluted and is unfit for human drinking. Infectious microorganisms may be present in human or animal waste. Wells and other drinking water sources can be contaminated by storm, water run-off from waste dump sites, landfills, roadways, farms and livestock operations, discharges from sewage treatment plants, or septic system discharges. In sub-urban areas like Evbotubu, access to clean, safe, and treated water is a major public health problem. In developing countries, access to both clean water and sanitation is limited. According to Fenwick (2006), two and a half billion people globally have no access to improved sanitation, and more than 1.5 million children die each year from diarrheal diseases. The isolation of *Bacillus* and *Staphylococcus* species from the dumpsites is in agreement with the studies by Williams and Hakam (2016). According to the researchers, both microorganisms produce enzymes known as Dnase and staphylokinase that could degrade wastes at dumpsites, and

convert it to useful materials.

Table 5 shows the antibiotic susceptibility pattern of bacterial and fungal isolates from the dumpsite soil, dumpsite leachate, well water, and borehole water samples. The result shows that ciprofloxacin was the most effective antibiotic based on the total percentage of isolates (45.37%) susceptible to the antibiotic. On the contrary, tetracycline was the least effective antibiotic that accounted for 12.04% of the isolates that was susceptible to the antibiotic. Interestingly, ciprofloxacin was most effective against *Staphylococcus aureus*, followed by *Pseudomonas aeruginosa*. It should be a great concern that the antibiotics were least effective against *Corynebacterium* spp. *Bacillus cereus*, and *Enterococcus* spp. The resistance demonstrated by these bacterial isolates against the antibiotics could be attributed to the production of enzymes, which inactivate or modify antibiotics. It is possible that the organisms could change its microbial cell membrane, modify the target site of the antibiotics, and metabolic pathways such that the antibiotics will have little effect on them (Odeyemi *et al.*, 2011). Therefore, continuous discharge of wastes in the dumpsites, and eventual infiltration into underground water present a high risk of spreading antibiotic resistant genes in the pathogens. Consequently, treatment of infections in humans caused by the pathogens harbouring antibiotic resistant genes becomes very difficult.

**Table 5.** Antibiotic susceptibility pattern of bacterial and fungal isolates from dumpsite soil, dumpsite leachate, well water and borehole water samples

| Microbial Isolates            | Number of Isolates | Percentage (%) of each bacterial isolate susceptible to antibiotics |        |       |       |       |        |       |       |       |       |       |       |
|-------------------------------|--------------------|---|--------|-------|-------|-------|--------|-------|-------|-------|-------|-------|-------|
|                               |                    | STR   | CIP    | ERY   | GEN   | OFL   | AMX    | COT   | PEF   | AUG   | NIT   | CAZ   | TET   |
| <i>Bacillus cereus</i>        | 11                 | 0.00  | 25.00  | -     | 25.00 | 20.00 | 25.00  | 25.00 | 33.33 | 0.00  | 50.00 | -     | 0.00  |
| <i>Enterococcus</i> spp.      | 6                  | -   | 33.33  | 25.00 | 33.33 | 0.00  | 0.00   | -     | -     | 0.00  | 25.00 | 33.33 | 25.00 |
| <i>Streptococcus</i> spp.     | 4                  | -   | -      | -     | 10.00 | 0.00  | 0.00   | 33.33 | 60.00 | 25.00 | 40.00 | 40.00 | 0.00  |
| <i>Escherichia coli</i>       | 6                  | -   | 40.00  | 0.00  | 0.00  | 25.00 | 33.33  | 0.00  | 33.33 | 66.67 | 0.00  | -     | 0.00  |
| <i>Corynebacterium</i> spp.   | 4                  | 25.00   | 10.00  | 0.00  | 0.00  | 33.33 | 0.00   | -     | -     | 25.00 | -     | 40.00 | -     |
| <i>Pseudomonas aeruginosa</i> | 4                  | 33.33   | 60.00  | 60.00 | 0.00  | 0.00  | -      | 40.00 | -     | -     | 0.00  | 25.00 | 0.00  |
| <i>Staphylococcus aureus</i>  | 5                  | 25.00   | 100.00 | 0.00  | -     | 60.00 | -      | 25.00 | 0.00  | 33.33 | 0.00  | 0.00  | 0.00  |
| <i>Fusarium</i> spp.          | 3                  | 50.00   | 40.00  | -     | 0.00  | -     | 100.00 | 0.00  | 25.00 | 0.00  | 33.33 | 25.00 | 25.00 |
| <i>Aspergillus</i> spp.       | 4                  | 0.00  | 60.00  | 25.00 | 25.00 | 0.00  | 33.33  | 25.00 | 66.67 | -     | 25.00 | 0.00  | 33.33 |
| <i>Mucor</i> spp.             | 2                  | 25.00   | 40.00  | -     | 25.00 | -     | 0.00   | 33.33 | 25.00 | -     | -     | 60.00 | 25.00 |
| Total                         | 49                 | 22.61   | 45.37  | 18.33 | 13.14 | 17.29 | 23.96  | 22.71 | 34.76 | 21.43 | 21.67 | 27.91 | 12.04 |

**Key:-** Represent culture plates that yielded no growth; 0.00 = Not susceptible; STR=Streptomycin; CIP=Ciprofloxacin; ERY= Erythromycin; GEN= Gentamicin; OFL=Ofloxacin; AMX= Amoxicillin; COT= Cotrimoxazole; PEF= Peflaxin; AUG= Augmentin; NIT=Nitrofurantoin;CAZ=Ceftazidimine; TET=Tetracycline

### Conclusion

The accumulation of waste in the dumpsite affected the physicochemical parameters, heavy metals, and minerals content of the dumpsite soil, well water, and borehole water to the extent that some of the results obtained were above the WHO permissible limits. Despite being located near a waste dumpsite, the quality parameters of borehole water is preferable than the well water. The isolation of pathogenic microorganisms from the soil and water sources, antibiotic resistance demonstrated by the isolates, calls for intensified efforts by the general public, and authorities to effectively manage open dumpsites in semi-urban communities to avoid outbreak of diseases.

### Conflicting Interests

Authors have declared that no competing interests exist

### REFERENCES

Abdus-Salam, N. (2009). Assessment of heavy metals pollution in dumpsites in Ilorin metropolis. *Ethiopian Journal of Environmental Studies and Management*, 2(2): 92-99.

Adejobi, O. S. and Olorunnimbe, R. O. (2012). Challenges of waste management and climate change in Nigeria: Lagos state metropolis experience. *African Journal of Scientific Research*, 7(1): 346 - 362.

Adejuwon, J. O. and Mbuk, C. J. (2011). Biological and physiochemical properties of shallow wells in Ikorodu town, Lagos Nigeria. *Journal of Geology and Mining Research*, 3: 161-168.

Adekola, P. O. and Ogundipe, A. A. (2017). An assessment of the state of environmental management in Nigerian capital cities. *Journal of Geography, Environment and Earth Science International*, 12(2): 1-13.

Adeyeba, O. A. and Akinbo, J. A. (2002). Pathogenic intestinal parasites and bacterial agents in solid wastes. *East African Medical Journal*, 79(11): 600–603.

Afolabi, T. A., Ogbunike, C. C., Ogunkunle, O. A., and Bamiro, F. O. (2012). Comparative assessment of the potable quality of water from industrial, urban and rural parts of Lagos, Nigeria. *Ife Journal of Science*, 14(2): 221–232

Agbeshie, A. A., Adjei, R., Anokye, J., and Banunle, A. (2020) Municipal waste dumpsite: impact on soil properties and heavy metal concentrations, Sunyani, Ghana. *Scientific African*, 8, 1–10.

Ahaotu, I., Wondikom, M. and Maduka, N. (2021). A preliminary study on the effect of storage temperatures on the population of *Bacillus cereus* in dry ginger powder. *Journal of Bioscience and Biotechnology Discovery*, 6(5): 53-57. <https://doi.org/10.31248/JBBD2021.159>

Ahmed, A. B. (2011). Insect vectors of pathogens in selected undisposed refuse dumps in Kaduna town, northern Nigeria. *Science World Journal*, 6(4): 21-26.

Akanchise, T., Boakye, S., Borquaye, L. S., Dodd, M., and Darko, G. (2020). Distribution of heavy metals in soils from abandoned dump sites in Kumasi, Ghana. *Scientific African*, 10: 1-12. <https://doi.org/10.1016/j.sciaf.2020.e00614>

Akhtar, M. M., Tang, Z., and Mohamadi, B. (2014). Contamination potential assessment of potable groundwater in Lahore,

Pakistan. *Polish Journal of Environmental Studies*, 23(6): 1905-1916.

Amano, K. O. A., Danso-Boateng, E., Adom, E., Nkansah, D. K., Amoamah, E. S., and Appiah-Danquah, E. (2021). Effect of waste landfill site on surface and ground water drinking quality. *Water and Environment Journal*, 35, 715–729.

Ano, A. O. and Okwunodudu, F. U (2008.). Effect of population and level of industrialization on underground water quality of Abia State, Nigeria. *African Journal of Biotechnology*, 7(4): 439–443.

AOAC (2000). Association of Official Analytical Chemists. Official Methods of Analysis 17 th Edn. The Association of Official Analytical Chemists, Arlington, Virginia, Gaithersburg, MD, USA.

APHA (2005). Standard methods for the examination of water and wastewater. American Public Health Association, American Water Works Association and Water Pollution Control 20th edition. Washington DC, USA pp. 5-17.5.

ASTM (2003). Standard methods for the examination if water and wastewater. American Standard for Testing and Materials, Washington DC.

Atikpo, E., Sada, S. O., Ihimekpen, M. O., and Okodugha, D. A. (2022). Solid waste generation rate for Evbotubu Community in Edo State of Nigeria. *FUW Trends in Science and Technology Journal*, 7(1): 419–423.

Bhalla, G., Swamee, P. K., Kumar, A., and Bansal, A. (2012). Assessment of groundwater quality near municipal solid waste landfill by an aggregate index method. *International Journal of Environmental Science*, 2(2): 1492–1503.

CLSI (2005). Performance standards for antimicrobial susceptibility testing, fifteenth informational supplement. Clinical and Laboratory Standard Institute, Wayne Pa M100 – S15 Vol. 25 no 1.

Didia, M. U. and Weje, I. I. (2020). Effects of refuse dump on ground water quality within the Rivers State University campus, Port Harcourt, Nigeria. *International Journal of Scientific and Research Publications*, 10(3): 697-706.

Enerijofi, K. E. and Ekhaise, F. O. (2019). Physicochemical and microbiological qualities of government approved solid waste dumpsites in Benin city. *Dutse Journal of Pure and Applied Sciences*, 5(2a): 12-22

Fenwick, A. (2006). Waterborne diseases - could they be consigned to history? *Science*, 313: 1077–1081.

Forsintus, N. O., Ikechukwu, N. E., Emenike, M. P., and Christiana, A. O. (2016). Water and waterborne diseases: a review. *International Journal of Tropical Disease*, 12(4): 1-14.

Fosu-Mensah, B. Y., Addae, E., Yirenya-Tawiah, D., Nyame, F., and Fantke, P. (2017). Heavy metals concentration and distribution in soils and vegetation at Korle Lagoon area in Accra, Ghana. *Cogent Environmental Science*, 3(1): 1–8.

George, M., Umadevi, A. G., Dharmalingam, P., Abraham, J. P., Rajagopalan, M., Balakrishnan, D. A., Harridasan, P. P., and Pillai, P. M. (2010). An investigation of quality of underground water at Eloor in Ernakulam district of Kerala, India. *E-Journal of Chemistry*, 7(3): 903–914.

Gupta, S. and Shukla, D. N. (2006). Physico-chemical analysis of sewage water and its effect on seed germination and seedling

- growth of *Sesamum indicum*. *J Nat-Ras. Development*, 1:5-19.
- Holt, J. G., Krieg, N. R., Sneath, P. H. A., Stanley, J. T., and Williams, S. T. (1994). *Bergey's Manual of Determinative Bacteriology* Williams and Wilkins Co. Baltimore Maryland 9th edition Lippincott, Williams and Wilkins, Philadelphia PA, 2000pp.
- Hossin, M. S., Matin, M. A., Islam, M. K., Rahman, M. M., Mukta, M. A., and Majumder, M. S. I. (2016). Water quality assessment of deep aquifer for drinking and irrigation purposes in selected coastal region of Bangladesh. *American Journal of Agricultural Science*, 3(6): 85-91.
- Ibiam, J. A., Oko, A. N., Amaechi, C. C., Apie, C. O., Nwali, U. I., and Isu, H. E. (2024). Effect of waste dumpsite on the surface and groundwater supplies using water quality index in Afikpo south local government, Ebonyi state. *World Journal of Advanced Research and Reviews*, 21(02): 1343-1352. DOI: <https://doi.org/10.30574/wjarr.2024.21.2.0547>
- Ichipi, E. B. and Senekane, M. F. (2023). An evaluation of the impact of illegal dumping of solid waste on public health in Nigeria: a case study of Lagos state. *International Journal of Environmental Research and Public Health*, 20:1-12. <https://doi.org/10.3390/ijerph20227069>
- Ishaku, H. T., Majid, M. R., Ajayi, A. P., and Haruna, A. (2011). Water supply dilemma in Nigerian rural communities: looking towards the sky for an answer. *Journal of Water Resource and Protection*, 3(8): 1-9. Doi: 10.4236/jwarp.2011.38069
- Joarder, S. D. (2000). Urban residential solid waste management in India: issues related to institutional arrangement. *Public Works Management Policy*, 4(4): 319-330.
- Jorgensen, J. H. and Turnidge, J. D. (2007). Antibacterial susceptibility tests: dilution and disk diffusion methods. In: Murray, P. R., Baron, E. J., Jorgensen, J. H., Landry, M. L., Pfaller, M. A. eds. *Manual of Clinical Microbiology* 9th ed. Washington DC. *American Society of Microbiology* p1152-1172.
- Khan, S. A., Din, Z. U., and Zubair, I. A. (2011). Levels of selected heavy metals in drinking water of Peshawar city. *International Journal of Science and Nature*, 2(3): 648-652.
- Kjeldsen, P., Barlaz, M.A., Rooker, A.P., Baun, A., Ledin, A., and Christensen, T. H. (2010). Present and long-term composition of MSW landfill leachate: a review. *Critical Reviews in Environmental Science and Technology*, 34(4): 297-336.
- Mahugija, J. A. M. (2018). Levels of heavy metals in drinking water, cosmetics and fruit juices from selected areas in Dar Es Salaam, Tanzania. *Tanzania Journal of Science*, 44(1): 1-11.
- Manetu, W. M. and Karanja, A. M. (2021). Waterborne disease risk factors and intervention practices: a review. *Open Access Library Journal*, 8: 1-11. DOI: 10.4236/oalib.1107401
- Manne, R., Kumaradoss, M. M. R. M., Iska, R. S. R., Devarajan, A., and Mekala, N. (2022). Water quality and risk assessment of copper content in drinking water stored in copper container. *Applied Water Science*, 12(27): 1-8. <https://doi.org/10.1007/s13201-021-01542-x>
- Moorberg, C. J. and Crouse, D. A. (2021). Soil Texture and Structure. *Soils Laboratory Manual*<https://kstatelibraries.pressbooks.pub/soilslabmanual/chapter/soil-texture-and-structure/>
- Nagarajan, R., Thirumalaisamy, S., and Lakshumanan, E. (2012). Impact of leachate on groundwater pollution due to non-engineered municipal solid waste landfill sites of Erode City, Tamil Nadu, India. *Iranian Journal of Environmental Health Science and Engineering*, 9(1):35.
- NIS (2007). Nigerian standard for drinking water. Nigerian Industrial Standards. Approved by SON Governing Council. Abuja/ Lagos HQ, pp 5
- Nkwunonwo, U. C., Odika, P. O., and Onyia, N. I. (2020). A review of the health implications of heavy metals in food chain in Nigeria. *The Scientific World Journal*, Volume 2020, Article ID 6594109, 11 pages. <https://doi.org/10.1155/2020/6594109>
- Nor Faiza, M. T., Hassan, N. A., Mohammad, F. R., Edre, M. A., and Rus, R. M. (2019). Solid waste: its implication for health and risk of vector borne. *Journal of Wastes and Biomass Management*, 1(2): 14-17. DOI: <https://doi.org/10.26480/jwbm.02.2019.14.17>
- Nwiniyi, O. C., Uyi, O., Awosanya, E. J., Oyeyemi, I., Ugbenyen, A. M., Muhammad, A., Alabi, O. A., Ekwunife, O. I., Adetunji, C. O., and Omoruyi, I. M. (2020). Review of drinking water quality in Nigeria: towards attaining sustainable development goal six. *Annals of Science and Technology*, 5(2): 58-77.
- Obasi, P. N. and Akudinobi, B. B. (2020). Potential health risk and levels of heavy metals in water resources of lead-zinc mining communities of Abakiliki, southeast Nigeria. *Applied Water Science*, 10 (184): 1-23. <https://doi.org/10.1007/s13201-020-01233-z>
- Obire, O., Nwaubeta, O., and Adué, S. B. N. (2002). Microbial Community of a waste dump site. *Journal of Applied Science and Environmental Management*, 6(1), 78-83.
- Odeyemi, A. T., Faweya, E. B., Agunbiade, O. R., and Ayeni, S. K. (2011) Bacteriological, mineral and radioactive contents of leachate samples from dumpsite of Ekiti State Destitute Centre in Ado – Ekiti. *Archives of Applied Science Research*, 3(4): 92-108.
- Odu, N. N., Omunakwe, A. L., and Millicent, M. (2020). Comparative assessment on the physicochemical water quality of wells and boreholes in two Rivers state communities, Nigeria. *International Journal of Research Studies in Microbiology and Biotechnology*, 6(3): 5-20. DOI: <https://doi.org/10.20431/2454-9428.0603002>
- Ogbeibu, A. E. (2005). *Biostatistics: A Practical Approach to Research and Data Handling*. Mindex Publishing Co. Ltd., Benin City. 264p
- Ogbu, S. U. and Ezeodili, W. O. (2021). Influence of rural-urban migration on waste management in the Enugu metropolis. *International Journal of Economics, Management, and Media Studies*, 3(2): 5-16. DOI: 10.36344/ccijemms.2021.v03i02.001
- Olukanni, D. O., Olujide, J. A., and Kehinde, E. O. (2017). Evaluation of the impact of dumpsite leachate on groundwater quality in a residential institution in Ota, Nigeria. *Covenant Journal of Engineering Technology*, 1(1): 1-16.
- Olukosi, O. M., Ameh, J. B., Abdullahi, I. O., and Whong, C. M. Z. (2016). Physicochemical quality of drinking water from various water sources of Kaduna state, Nigeria. *Bayero Journal of Pure and Applied Sciences*, 9(2): 141-144. <http://dx.doi.org/10.4314/bajopas.v9i2.26>
- Onwughara, I. N., Nnorom, I. C., and Kanno, O. C. (2010). Issues of roadside disposal habit of municipal solid waste, environmental impacts and implementation of sound management practices in the developing country Nigeria. *International Journal of Environmental Science and Development*, 1(5): 409 - 418.



- Oviasogie, F. E., Ajuzie, C. U., and Ighodaro, U. G. (2010). Bacterial analysis of soil from waste dump site. *Archives of Applied Science Research*, 2(5): 161–167.
- Oyekan, T. K. and Sulyman, A. O. (2013). Health impact assessment of community-based solid waste management facilities in Ilorin West Local Government Area Kwara State, Nigeria. *Journal of Geography and Regional Planning*, 8(6): 26–36.
- Oyeleye, O. O. (2013). Challenges of urbanization and urban growth in Nigeria. *American Journal of Sustainable Cities and Society*, 2(1): 79-95.
- Pandey, A., D'souza, N., and Satpute, A. (2021). Total organic carbon analyzer: emerging technique for quality assessment of potability of drinking water. *Indian Journal of Advances in Chemical Science*, 9(1): 47-50.
- Salaimon, A. M., Akinwotu, O. O., and Amoo, O. T. (2015). Resistance of bacteria isolated from Awotan dumpsite leechate to heavy metals and selected antibiotics. *International Journal of Research in Pharmacy and Bioscience*, 2(9): 8–17.
- Sargar, G. S., Thakare, B. G., and Hasekar, A. S. S. (2024). A review of permissible limits of physicochemical parameters drinking water. *International Journal of Research and Analytical Reviews*, 11(1): 180-185.
- Tsuneo, W. (2010). Pictorial atlas of soil and seed fungi. Morphologies of cultural fungi and key to species. 3<sup>rd</sup> edition, CRC Press.
- Tukura, E. D., Ojeh, V. N., Philip, A. H., and Ayuba, A. (2018). Assessing the potential health effect of solid waste dump site located close to residential areas in Jalingo, Taraba state using geospatial techniques. *World News of Natural Sciences*, 20: 160-175.
- Udofia, U. U., Udiba, U. U., Udofia, L. E., Ezike, N. N., and Udiba, S. U. (2016). Assessment of the impact of solid waste dumps on ground water quality, Calabar municipality, Nigeria. *Journal of Advance Research in Pharmacy and Biological Science*, 2(4): 10-21. DOI: <https://doi.org/10.53555/nnpbs.v2i4.705>
- Ugbebor, J. N. and Ntesat, B. (2019). Investigation of borehole water contaminant profile at Igwuruta solid waste dumpsite, Rivers State. *Nigerian Journal of Technology*, 38(2): 532-539.
- Ugwoha, E., Udeh, N. U., and Kpaniku, D. (2023). Impact of dumpsite on neighbouring boreholes in Igbogo community of Rivers state, Nigeria. *Asian Journal of Geological Research*, 6(1): 1-11.
- Ukpong, E. C., Udo, E. A., and Umoh, I. C. (2015). Characterization of materials from Aba waste dump sites. *International Journal of Engineering and Applied Sciences*, 6(3): 1-10.
- WHO (2010). Guidelines for drinking water quality. Guidelines for drinking-water quality, incorporating 1st and 2nd Addenda, Vol. 1, Recommendations, 3rd ed.; WHO: Geneva, Switzerland.
- Williams, J. O. and Hakam, K. (2016). Microorganisms associated with dumpsites in Port Harcourt metropolis, Nigeria. *Journal of Ecology and the Natural Environment*, 8(2): 9–12.
- Yahaya, T., Abdulganiyu, Y., Gulumbe, B. H., Oladele, E., Anyebe, D., and Shemishere, U. (2022a). Heavy metals and microorganisms in borehole water around the Olusosun dumpsite in Lagos, Nigeria: occurrence and health risk assessment. *Avicenna Journal of Environmental Health Engineering*, 9(2): 1-6. DOI: 10.34172/ajehe.2022.09
- Yahaya, T. O., Abdulganiyu, Y., Ologe, O., Muhammad, Z. K., Ibrahim, Y. Y., Haruna, A. J., and Adeniyi, T. J. (2022b). Quality and safety assessment of borehole water around Simpson transfer loading station in Lagos, southwest Nigeria. *Dutse Journal of Pure and Applied Sciences*, 8(4): 20-35. <https://dx.doi.org/10.4314/dujopas.v8i4a.3>